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Review of Wastewater Treatment Technologies in View of their Application in the DR Congo Mining Industry

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ABSTRACT

The world is currently facing the water crisis brought about by the deterioration in its quality and the difficult access to its resources due to ever-growing demand in large-scale consumer sectors such as agriculture, industry and tourism. Consequently, increasing pressure on water resources is experienced in the world as the outcome of the strong population growth, continued urbanization and rapid industrialization, together with numerous discharges of polluted waters into watercourses. Strategies have been developed throughout the world for improving both the quality and access to water in a sufficient quantity. Thus, different technologies, both established and emerging ones, are presently utilized in view to treat wastewaters, including those of the mining origin. The present work reviews the different uses of water in the mining industry and identifies different sources of polluted waters. It analyses wastewater treatment techniques in view of their application specifically to the treatment of wastewaters from the DR Congo mining industry. As such, this work identifies itself as part of an effort to contribute to the improvement of wastewater management practices in the DR Congo mining industry as well as improved access and sustainable use of water resources.

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1. Introduction

Fresh water is a scarce natural resource very essential to life and any productive activity [1-9]. Water worldwide availability or accessibility (Figure 1) is not uniform [10] given that some regions have significant water resources, while others, such as in the Northern region of China [11], Chile and the WANA region endure water scarcity and drought [2, 6, 10-14]. The WANA region includes countries in North Africa (Algeria, Morocco, Tunisia, Libya, Egypt, Eritrea, Ethiopia, and Sudan), West Asia (Turkey, Cyprus, Iraq, Jordan, Lebanon, and Syria) as well as all the Middle Eastern countries in the Arabian Peninsula and Iran.

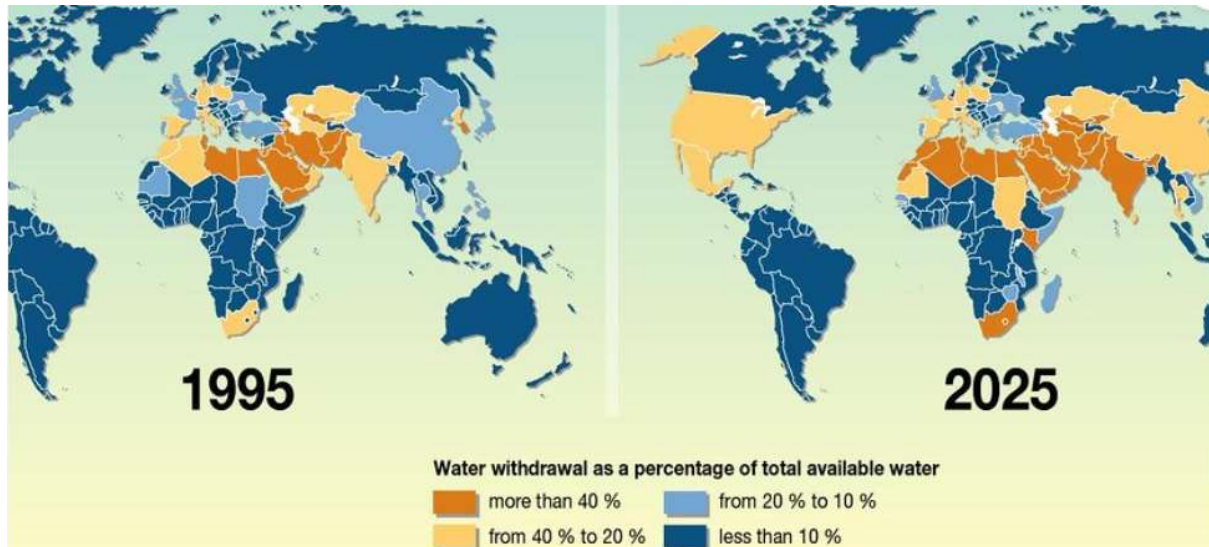


Figure 1: Water withdrawal as a function of total water available.

About 1.2 billion people do not have access to water around the world [6]. While sustainable development advocates access to good water and in sufficient amount for all [15, 16], 2.4 to 2.6 billion people are deprived of this vital resource [2]. Consequently, those people do not enjoy a minimum of health conditions and die from waterborne diseases [2, 17, 18]. This situation has affected more than 66% of the Asian population [19] of which the majority belongs to the WANA region, which accounts for a major proportion of the world's driest areas [2].

The world is currently experiencing a water crisis from the point of view of both quality deterioration and the lack of access to sufficient quantity [2, 18]. According to [16], Water scarcity means the lack of sufficient water, or not having access to safe water supplies and this concept is clearly different from water stress. This refers to a lack to meet human and ecological demand for water during a certain period; a situation experienced by a large proportion of the world's population [20]. Indeed, there is increasing pressure on water resources induced by population strong growth and continuous urbanization, rapid industrialization as well as the intensification of food production [6, 8, 10, 21]. Another issue to be mentioned, in relationship with the water crisis, is the increasing number of discharges into waterways of polluted waters [4, 5, 22-25] in the majority of countries [2, 18]. The water crisis has resulted in over-exploitation of aquifers, rampant pollution of the environment [26] and shortages that have alienated the universal human right to have good quality water in sufficient quantity at one's disposal [2, 5, 9, 17].

[9] argue that the adoption of water recycling practices, implementation of wastewater treatment processes and use of remediation systems has become inescapable routes for increasing access to water [6, 27]. Thus, strategies have been developed throughout the world in view of improving access to water [6]. Different technologies, both established and emerging ones, are utilized in view to treat raw water and to produce safe water [9]. These technologies are utilized for treating wastewaters of different origins, including those from the mining industry and enable recovering safe water that is either used in other sectors or released to watercourse [1, 12, 28-39].

In 2003, for instance, global wastewater production was estimated at approximately 1,500 km³ [40]. In emerging countries such as China [41], the volume of wastewater of the mining origin was estimated at 1,028 mega tonnes, an amount that accounted for about 6% of the total volume of industrial wastewater (Table 1).

Table 1: Industrial wastewater production in China

Industrial Sector	Number of Enterprises	Total Volume (Mega Tonne)	Standard Volume (Mega Tonne)	Proportion (%)
Food and brewing	11,672	2,546	2,257	11.0
Chemical	9,326	3,240	3,040	14.8
NFS Manufacturing	21,402	403	367	1.8
Metal manufacturing	5,995	333	316	1.5
Foundry	2,408	318	299	1.5
Metallurgical	6,086	1,887	1,825	8.5
Mining	2,314	594	533	3.0
Paper	5,818	4,246	3,038	18.7
Pharmaceutical	2,605	429	404	2.0
Textile	7,644	2,252	2,047	10.0
NFM processing	1604	434	384	1.9
Total	106,434	22,076	20,471	100

NFM: non ferrous metal, **SNF:** Non ferrous substances.

Regarding the management of mine wastewater in Europe, statistics reveal that in Germany, for instance, about 12% of the total volume is handled. Surface water receives 5,750 million m³ of wastewater. Industrial wastewater treatment is recommended owing to its high pollution capacity [42, 43]. Indeed, the mismanagement of one litre of wastewater can result in the pollution of 8 litres of fresh water [40]. Mining wastewater treatment is done using different physical, chemical or biological technologies [22, 44]. The treatment of wastewater consists of a combination of physical, chemical and/or biological processes and operations enabling the removal of contaminants, including colloids, organic and inorganic matters, nutrients as well as soluble matters (metals, organics, etc.) [5, 26, 45- 47].

The present work reviews the different uses of water in the mining industry and identifies different sources of polluted water. It analyses wastewater treatment techniques in view of their application specifically to the treatment of wastewaters from the DR Congo mining industry. As such, this work is part of an effort to contribute to the improvement of wastewater management practices in the DR Congo mining industry as well as improved access and sustainable use of water resources.

2. Water Uses in the Mining Industry

Water is vital for any industrial activity [18, 48] because it is used as a solvent, as the main ingredient of industrial processes, as a reagent, as cleaning liquid, as reaction and transport medium and as a medium for reactants, as heat exchange fluid or as a source of electrical energy [49-51]. With about 25% of global demand for water, industry ranks second only to agriculture in terms of water consumption [18, 50].

Water availability is thus one of the key variables in the design and materialization of a mining project [48]. Indeed, water is used at all stages (Figure 2) of a mine's life, that is, during the exploration phase of the ore deposits as well as during the mining and processing of ores. Water is used in dust suppression, worker camp supply, irrigation of surrounding lands and rehabilitation of mine sites [52].

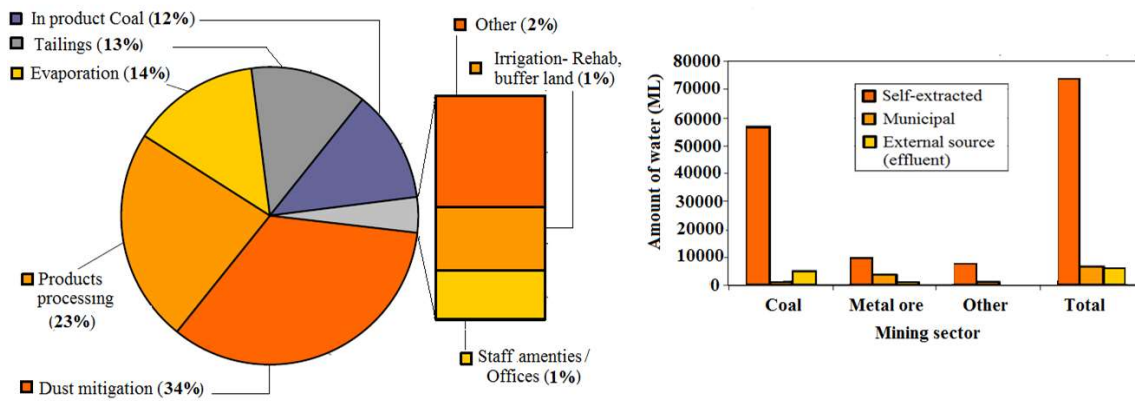


Figure 2: Water use at a mine in Queensland [15].

Table 2 below describes water consumption at some base metal mines in Australia [53].

Table 2: Water consumption in some Queensland mines

Mine	Mined Ore	Daily Water Consumption (Mega Litre)	Annual Water Consumption (Mega Litre)
A	Copper-gold	13-17.8	4,750-6,500
B	Lead-Zinc-Silver	6	2,200
C	Copper-gold	5.2	2,000
D	Copper-gold	1.2	450
E	Lead-Zinc-Silver	26-52	9,500-19,000

For instance, from the mining of ore until the extraction process of copper as metal, water is used in the achievement of operations described in Figure 3.

2.1. Water Supply to the Workers' Camp

The proportion of water intended for feeding a mining camp, particularly for the usual domestic needs and environmental sanitation, is by far negligible compared to that requested by mining operations.

2.2. Water Used in Mining Operations

The proportion of water consumed by mining operations is estimated at about 15% of the total water volume consumed at a mine site. The largest water consumption at an open-pit mine site is due to dust suppression. It depends on the morphology of the terrain and varies according to the seasons and in arid areas, it may be more important [12]. For mining sites located in regions with high rainfall, dust suppression is carried out naturally resulting in lowering of the overall amount of consumed water.

In an underground mine, the proportion of water required for dust suppression is small [12]. Depending on the conditions of a mine site, that proportion varies between 0 and 15% of the total volume of water used for all mining operations. This water comes from the pumping of groundwater or from the collection of runoff waters [12, 54].

2.3. Water Supply to an Ore Concentrator

It represents the largest volume of water consumed at a mine site [12, 54]. Indeed, grinding and beneficiation of ores through flotation or gravity separation usually consumes large volumes of water. Copper ores, for

instance, can be mined as sulphides or oxides and their beneficiation through flotation is the most water-consuming operation given that it is conducted using pulps (Figure 4) containing 25 to 45% of solids [12].

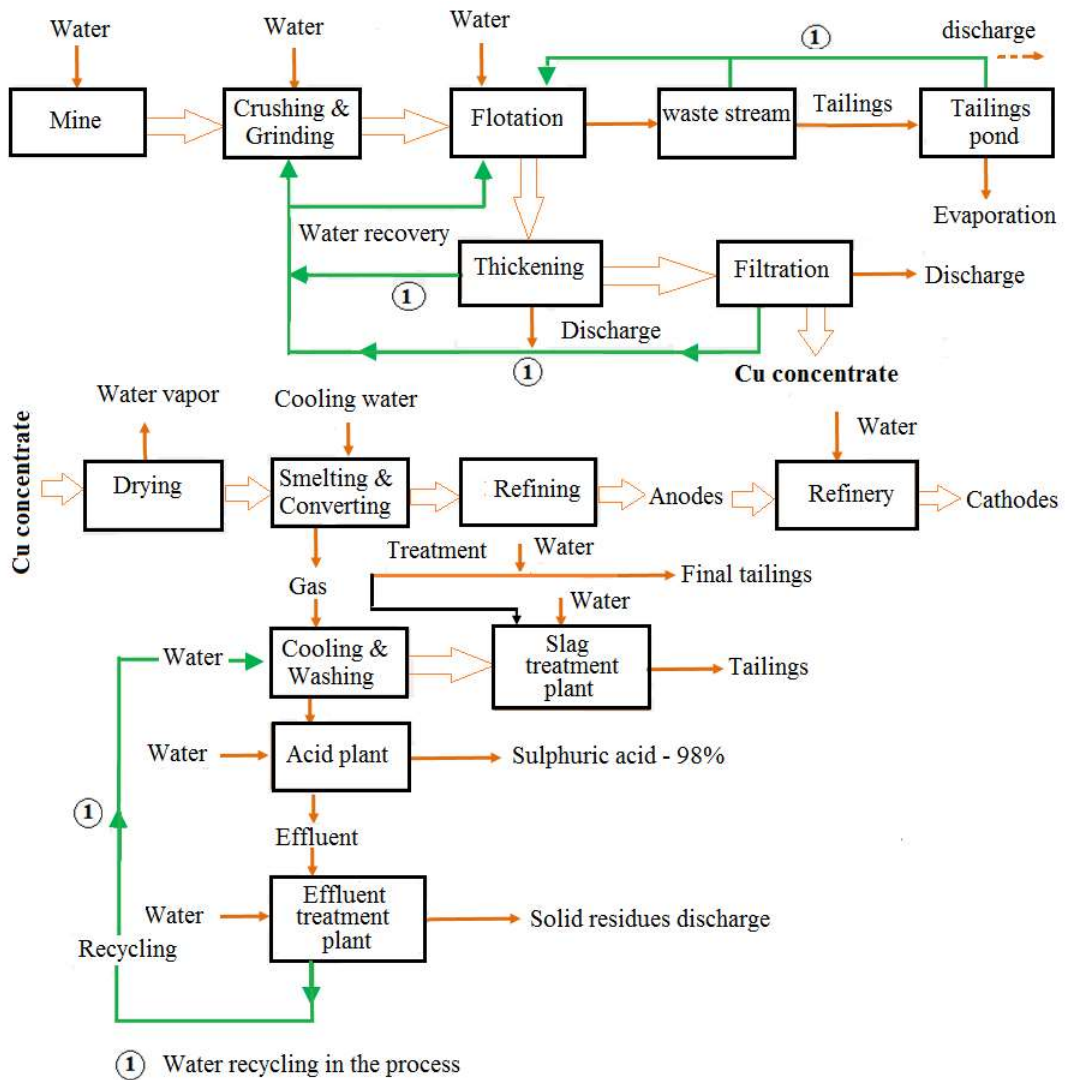


Figure 3: Water consumption during the pyrometallurgical production of copper (modified from [12]).

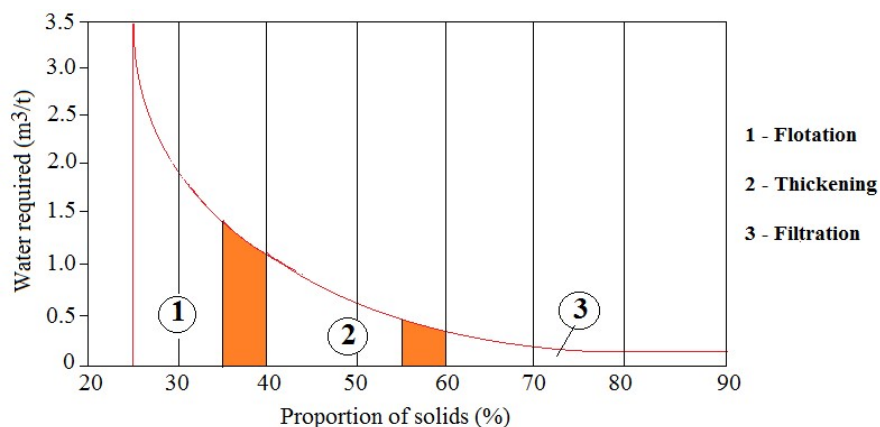


Figure 4: Water consumption during the flotation beneficiation of copper ores (After [12]).

Flotation is reputed a water-consuming mineral processing operation that enables economically separating valuable minerals from the gangue ones [33, 55]. Water flow rates in the range of 300 to 500 m³/h are utilized in the flotation of copper-cobalt oxidized ores mined in the Katanga region of the DR Congo [56, 57]. The variation in the proportion of solids in the pulp (Figure 4) is revealing that most of the feed water (80-90%) finds one's in the discharges from the ore flotation. Settling and filtration operations of the final concentrate enable recovering wastewater from the flotation of ores [12, 56], with the recovered wastewater either mixed with flotation tailings or returned to the process [12, 32, 33].

In copper ores flotation, for instance, water consumption ranges from 1.5 to 3 m³/t [12, 54]. Milling and flotation of Pb-Zn polymetallic ores require water flow rates in the order of 6000 L/h [58]. As for the production of one tonne of metallic copper, for instance, it requires approximately 13.6 m³ of water, that is, from the mining of ores until their flotation beneficiation.

The water consumption of a concentrator is estimated at 0.75 m³ per tonne of ore. Maximizing recycling and prevention of water leakage, together with losses due to evaporation, can help optimize water consumption up to 0.34 m³ per tonne of ore, as is the case in some ore concentrators operated in Chile and Australia [54]. The life cycle analysis showed that water consumption in terms of tonnes for the production of various base metals ranges from 3 to 250 m³. Its extreme values have been observed in gold and steel manufacturing [59].

Table 3 illustrates the water consumption during the production of some base metals, that is, from the mine until the ores beneficiation through flotation.

Table 3: Water consumption from the mine to the base metal

Metal	Water Consumption (m ³ Per Tonne)
Copper	7.8
Nickel	0.93
Lead	0.64
Zinc	0.64
Iron/Steel	0.21
Titanium	5.15
Aluminium	0.3

2.4. Water Used to Transfer Ores and Concentrates

During ores processing, minerals are moved through pipes in the form of pulps from one machine to another [48]. The same applies to concentrates when they have to pass from one enrichment stage to another or be sent to the thickening and to dehydration by vacuum filtration in view obtaining a cake with the moisture in the range 10 to 15%, that is, the final concentrate.

Some concentrators are designed so that the wet grinding section is situated away from the ores flotation section. This implies transferring from one processing section to another of large tonnages of solid materials in the form of pulps. Sometimes the settling and filtration sections of the final concentrate are located in a seaport in view facilitating the export of the concentrates to an electrical smelter. This is the case at the Escondida mine or to that of the Pelambres located in Chile [12]. The proportion of water dedicated to this operation accounts for 4 to 6% of the total volume consumed by an ores concentrator. The Alumbrera Mine (Argentina) possesses the longest pipeline (316 km) in the world for transferring ores concentrates (Cu-Ag) until to the settling and filtration plant using water pumping [60].

2.5. Water Supply to a Foundry Cooling System

In an electrical foundry, for instance, the production of copper blister or copper anodes is done using the smelting-conversion process conducted inside specialized metallurgical reactors [12]. The feed is often composed of dry sulphide concentrates containing 0.2% humidity. Their metallurgical conversion into mattes requires oxygen of which the production consumes large amounts of water [12]. An amount of 7.8 m³ of water is required for the smelting process in view of obtaining one tonne of copper metal and an additional 0.6 m³ of water is required for refining operations of the obtained metal [59].

In an electric foundry, water is also utilized as a coolant for furnaces, in the granulation process of slag and metal alloys as well as during the cleaning of the smelting furnace exhaust gases. This is the reason why several smelters are located near the watercourses [56]. Indeed, they collect their feed water upstream and discharge the wastewater downstream in rivers used as spillways [57, 61]. The Big Hill Smelter in Lubumbashi (the "Société du Terril de Lubumbashi" in French), operated in Katanga region since 2002 up to 2017 and again until nowadays, illustrates this practice with the Lubumbashi River used as the source of the electric arc furnace cooling water and as a spillway of the process wastewater. This smelter had produced copper and cobalt alloys through reprocessing of slag from the pyrometallurgy extraction of copper [61], with the cooling system consuming 70-75 m³ of water per hour abstracted from the Lubumbashi River. In Chile, studies have shown that electric smelters of copper consume on average 3.6 m³ of water per tonne of metal for the functioning of their cooling systems [12].

2.6. Water Supply to a Metal Refining Plant

Metal anodes of copper produced by an electric foundry are refined through electrolysis (60°C) in view removing impurities so that their concentrations become as low as 0.1-0.3% [12]. During metal refining, water consumption is due to evaporation losses as well as, the bleeding of electrolytes loaded with impurities (As, Sb, Cl, etc.) in view of diminishing their concentrations in the processing circuit [62].

3. Sources of Wastewater of Mining Origin

Throughout their lives, mines discharge to the environment various types of wastewater [24, 63], that is, since the exploration stage of ore deposits until the extraction of metals of interest. The released wastewater contains a variety of pollutants [24, 33, 63, 64] responsible for the deterioration of the quality of groundwater and surface water (Table 4).

Major sources of wastewater of mining origin are described below:

3.1. Acid Mine Drainage (AMD)

The main source of polluted waters in relationship with mining activities is acid mine drainage (AMD) [11, 44, 65- 68]. The AMD is an acid-generating process that occurs following exposure to air and water of reactive sulphides such as flotation tailings or mining wastes [65, 66, 68, 69]. It is the biggest environmental issue mining companies are confronted with, especially during the rehabilitation of abandoned mine sites or after closure [65-67, 70].

In 1989, it had been reported that about 19,300 km of rivers, 72,000 ha of lakes and water reserves were negatively impacted by liquid effluents from the mining industry [67]. According to [66], AMD acidifies the receiving environment through the liberation of sulphuric acid, and the maintaining in aqueous solutions of highly reactive and toxic metals [65, 68]. Indeed, it is difficult to stop an AMD and it cannot do it on its own. The generation of persistent acids observed in mining areas, around 1700 in Norway, is perfectly illustrating this phenomenon [67, 71]. In addition, the costs of containment of pollution due to AMD are such that they can swallow up over the long term the economic gain from an initially profitable mining operation. Indeed, the neutralization of water polluted by an AMD at the abandoned site of mine Equity (Houston-British Columbia), for instance, costs annually more than 1.2 million USD to the mining company. This acid generation process is expected to continue for 150,000 years [71].

Table 4: Physicochemical quality of liquid effluents from ore flotation

Parameter (Unit)	Unfiltered Liquid Effluent	Liquid Effluent from a DRC Concentrator
pH	8.0	10.11
Turbidity (NTU)	7.5	NA
Electric conductivity ($\mu\text{S}/\text{cm}$)	171	4,290
Colour (TCU)	NA	550
Saltiness	NA	1.9
Matter in suspension (mg/L)	NA	674
Copper (mg/L)	0.03	0.38
Cobalt (mg/L)	NA	0.24
Manganese (mg/L)	NA	0.40
Zinc (mg/L)	0.32	NA
Lead (mg/L)	0.013	0.09
Nickel (mg/L)	NA	0.03
Arsenic (mg/L)	0.011	NA
Iron (mg/L)	0.23	0.34
Cadmium (mg/L)	0.008	0.01
Mercury (ppb)	0.66	NA
Sulphate ions (mg/L)	NA	364

NA: not available

Acid generation is a process that therefore appears under certain natural conditions. Its exacerbation is due to mining activities, that is, the excavation and milling of sulphide minerals and their exposure to air and water outside their natural environment [65, 66]. The magnitude of AMD is determined by the sulphur content of reactive minerals contained in flotation tailings or mining wastes [71] and it is enhanced by the presence of *Thiobacillus ferrooxidans* bacteria [66]. AMD results from natural oxidation of sulphides or pyrite-rich tailings [65, 66, 70], with a generation of sulphuric acid based on the weathering reactions shown in Table 5 [69].

Table 5: Oxidation paths of sulfide minerals

Oxygen Path	
Pyrite	$\text{FeS}_2 + \text{H}_2\text{O} + 7/2\text{O}_2 \rightarrow \text{Fe}^{2+} + 2\text{SO}_4^{2-} + 2\text{H}^+$
Chalcopyrite	$\text{CuFeS}_2 + 2\text{O}_2 \rightarrow \text{Fe}^{2+} + \text{Cu}^{2+} + 2\text{SO}_4^{2-}$
Sphalerite	$\text{ZnS} + 2\text{O}_2 \rightarrow \text{Zn}^{2+} + \text{SO}_4^{2-}$
Galena	$\text{PbS} + 2\text{O}_2 \rightarrow \text{Pb}^{2+} + \text{SO}_4^{2-}$
Arsenopyrite	$\text{FeAsS} + 3.25\text{O}_2 + 1.5\text{H}_2\text{O} \rightarrow \text{Fe}^{2+} + \text{HAsO}_4^{2-} + \text{SO}_4^{2-} + 2\text{H}^+$
Ferric iron path	
Pyrite	$\text{FeS}_2 + 14\text{Fe}^{3+} + 8\text{H}_2\text{O} \rightarrow 15\text{Fe}^{2+} + 2\text{SO}_4^{2-} + 16\text{H}^+$
Chalcopyrite	$\text{CuFeS}_2 + 16\text{Fe}^{3+} + 8\text{H}_2\text{O} \rightarrow 17\text{Fe}^{2+} + \text{Cu}^{2+} + 2\text{SO}_4^{2-} + 16\text{H}^+$
Sphalerite	$\text{ZnS} + 8\text{Fe}^{3+} + 4\text{H}_2\text{O} \rightarrow 8\text{Fe}^{2+} + \text{Zn}^{2+} + \text{SO}_4^{2-} + 8\text{H}^+$
Galena	$\text{PbS} + 8\text{Fe}^{3+} + 4\text{H}_2\text{O} \rightarrow 8\text{Fe}^{2+} + \text{Pb}^{2+} + \text{SO}_4^{2-} + 8\text{H}^+$

These reactions are the sources of mineralogical and geochemical alterations of sulphide minerals leading to the sulphuric acid leaching of oxidation products, with the release of metal elements to the environment [65, 68, 72]. The absence of neutralizing substances (Table 6) in sulphide-rich tailings promotes the establishment and spreading of an AMD [69].

Table 6: Compounds inducing chemical reaction involved in the consumption of acid

Neutralizing Compound	Reaction Leading to the Consumption of Acid Generated During the Oxidation of Sulfides
Calcite	$CaCO_3 + 2H^+ \rightarrow Ca^{2+} + CO_2 + H_2O$
Chlorite	$(Mg_{4.5}Fe_{0.2}^{II}Fe_{0.2}^{III}Al)AlSi_3O_{10}(OH)_8 + 16H^+ \rightarrow 4.5Mg^{2+} + 0.2Fe^{3+} + 2Al^{3+} + 3SiO_2 + 12H_2O$
Plagioclase	$Na_{0.75}Ca_{0.25}Al_{1.25}Si_{2.75}O_8 + 5H^+ \rightarrow 0.75Na^+ + 0.25Ca^{2+} + 1.25Al^{3+} + 2.75SiO_2 + 2.5H_2O$
Potassium feldspar	$KAlSi_3O_8 + H^+ + 4H_2O \rightarrow K^+ + 3H_4SiO_4 + Al(OH)_3$

Neutralizing substances, therefore, block the formation and spread of AMD via consumption of the acid generated due to the weathering of sulphide minerals. Neutralizing substances act through the increase in the pH causing the precipitation of metal elements contained in the water of which the quality has been affected by AMD, which process allows the containment of the pollution. This is the reason why the control of pH is very important before releasing to the environment wastewater from mining origin [73]. In the Cities of Kipushi and Kinsenda (DR Congo), where large amounts of sulphide tailings have been exposed to air and water since many years, the risk of AMD spreading is minimized given that the storage sites are located on dolomitic geological formations known as the Kakontwe limestone [56, 57, 74].

AMD is therefore a source of contaminated and aggressive waters (rich in sulphates and carbonates) that are emitted owing to its dissolving capacity enhanced by acidity and bacteria. Thus, it is a natural process capable of negatively impacting water quality in a variety of ways [68, 71]. Indeed, AMD usually lowers the pH of water, resulting in increased mobility and the build-up of metal elements up to concentrations qualified as toxic toward any form of life (Table 7). Copper and arsenic concentrations observed, for instance, at Buck Creek (Ohio-USA) were 750 and 20 times higher than their ordinary values [71], with the water rendered unsuitable for any use [75].

Table 7: Water from an abandoned copper mine polluted by the AMD

Parameter	pH	Fe (mg/L)	SO ₄ ²⁻ (mg/L)	Cu (µg/L)	Zn (µg/L)	As (µg/L)
Value	3,83	542	3642	1880	9599	108

AMD had caused significant environmental damages such as the deforestation of large tracts of land and pollution of both surface and groundwater [75, 76].

3.2. Mine Extraction Operations

At the mining stage, water is used in drilling as a lubricant, in washing operations of mining equipment and in many other applications [59]. The excavations were done on a given mining site usually result in the release of highly mineralized waters containing lubricant oils due to leaks happening on mining machinery and equipment. Those waters contain also fine solid particles, large amounts of cations and anions such as sulphate and chloride ions. Nitrate and ammonium ions are also present in those waters type mainly as the outcome of mining operations. The best management of such mineralized waters is essential in order to prevent contamination of both ground water and surface one.

3.3. Mineral Processing and Tailings Management

Metallurgical processes release large amounts of acidic or alkaline-type polluted waters in which high concentrations of trace metals, cyanides, sulphates, etc. can be present. The ores flotation is an important source of polluted waters. Indeed, different reagents are utilized to separate useful minerals from the gangue. In parallel,

the metal elements are liberated into the process water by the surface dissolution of minerals subjected to flotation. Unconsumed reagents, together with their degradation products, are also transferred to the wastewater generated due to the ores flotation [77].

The flotation of sulphide ores using xanthates as collectors together with the grinding and beneficiation of oxidized minerals of base metals through flotation, of course after their sulphidization, result in the formation of thiosalts in the process water [78, 79]. The generation of thiosalts during the flotation of a sulphide ore, containing pyrite, for instance, takes place using the following chemical reactions' sequence:



Thiosulphate ions generated by the grinding and beneficiation of ores through flotation are converted into polythionate ions and further on, into the sulphate ions as the oxidation process end-products (Figure 5). Thiosalts ($S_xO_y^{2-}$) is partially converted into sulphur-bearing more oxidized compounds considered as the products of sulphide minerals' alteration.

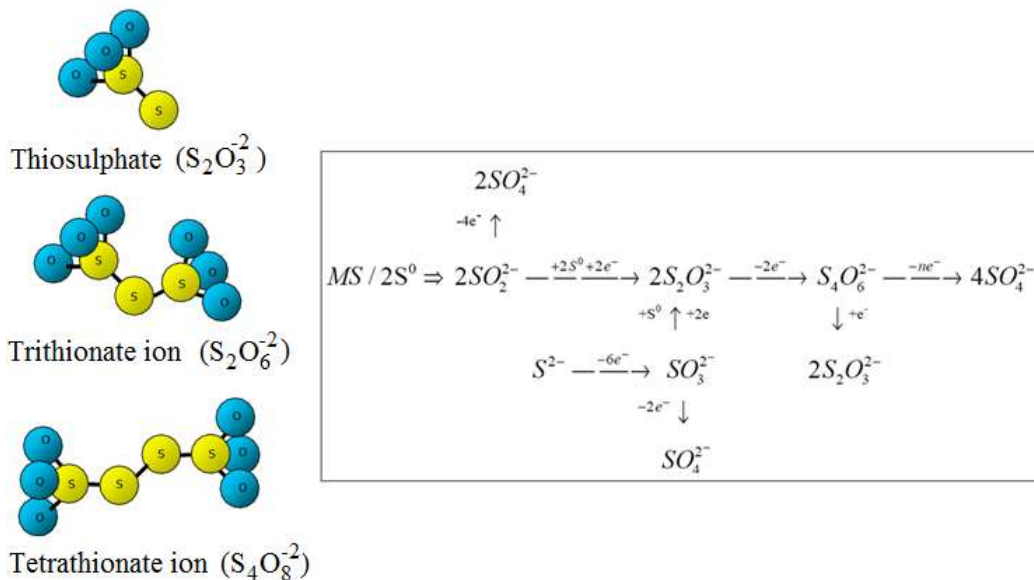


Figure 5: Thiosel generation during the flotation of sulphide ores (After [79]).

Apart from the unconsumed flotation reagents and metal elements of which the formation has been highlighted above, the wastewater from the flotation of either sulphide or oxidized minerals also contains the thiosalts generated due to oxidation of flotation reagents (Na_2S , $NaHS$ and xanthates) by reaction with air introduced into the pulp. As can be seen from the reactions (5) to (7), the conversion of thiosulphate ions and polythionate ions into sulphate ion is bringing about the acidification of the process water. This process renders unsuitable water for recycling in ores flotation [55, 79, 80]:



In addition, the degradation kinetics of thiosalts, during the storage in tailings ponds of liquid effluents from ores flotation, is sometimes so slowed that some of them are not converted into sulphate ions [78]. Subsequent oxidation of thiosalts into sulphate ions, through bacteria-induced processes (Table 8), leads to the generation of acid in watercourses utilized as spillways for flotation wastewater [78, 79, 81].

Table 8: Bacteria involved in thiosalts oxidation

Name	Substrates Utilized	pH
<i>Thiobacillus thiooxidans</i>	Sulphur, tetrathionate, thiosulphate.	Optimum 2.0 – 3.0 Range 1.0 – 6.0
<i>Thiobacillus thioparius</i>	Sulphur, trithionate, Tetrathionate Dithionate, thiosulphate.	Optimum near neutrality
<i>Thiobacillus neapolitanus</i>	Sulphur, thiosulphate.	Optimum 7.0 Range 4.5 – 7.8
<i>Thiobacillus novellus</i>	Thiosulphate, organics	Optimum 7.0 Range 5 – 9.5
<i>Thiobacillus detrificans</i>	Sulphur, Thiosulphate, Dithionate, Tetrathionate, sulphide.	Range 6 – 8

The bio-oxidation of thiosalts can therefore provide elemental sulphur, highly oxidized polythionates, or lead to sulphate ions formation, depending on the environment that prevails inside tailings ponds. Under well-controlled conditions, the final product from thiosalts oxidation is the sulphate ion as per the reaction chemical (7). The oxidation of thiosalts increases the dissolving capacity of water, a process capable of bringing about secondary pollution through the redissolution of metal elements initially trapped in waterway sediments. Liquid effluents from the ores flotation can thus endanger the flora, fauna and human health through the pollution of receiving media [26, 79].

4. Treatment Technologies for Wastewaters of Mining and Metallurgy Origin

4.1. Physical Treatment

This treatment utilizes several techniques adapted to the decontamination of different types of wastewater. A special focus is put on the description of techniques for the treatment of polluted water of mining origins.

4.1.1. Water Treatment Using Membrane Technologies

Membrane technologies are used in the treatment of water containing pollutants such as polycyclic aromatic hydrocarbons and other pollutants [10]. During the treatment of polluted waters, these technologies utilize mainly the size exclusion as a mean to remove pollutants apart from other mechanisms such as hydrophobic effects and adsorption on the membrane surface [10]. Membrane technologies comprise microfiltration, ultrafiltration, nanofiltration and reverse osmosis [10, 82]. The latter is a technique [37, 63, 83] used for water purification and based on membrane filtration under high pressure (2000 to 8000 k Pa). Reverse osmosis, which accounts for more than 30 years in the production of drinking water, is now increasingly used in wastewater treatment [9, 84]. Indeed, with the strengthening of environmental regulation in both developed countries and developing ones, membrane filtration (Figure 6) has become since 1990 a more attractive option in the production of recyclable water from industrial wastewater [1, 9, 37].

In Chile, reverse osmosis is utilized in the treatment of recyclable water in the copper industry [85]. It is also implemented in the treatment of water for the recharge of overexploited aquifers or wells threatened by drying up. However, reverse osmosis, as it is the case of other membrane technologies, remains a very expensive water treatment technique (€0.76-2.12/m³/year) owing to its high energy consumption and the scaling of nano-filters resulting in decreased performances over time [6, 10, 37, 86]. This water treatment technology is thus accessible only in the world's wealthy regions, such as Israel, as well as Chile and Australia, which are areas very confronted

with severe drought [9]. However, according to [21], treatment costs have significantly decreased over the last few decades owing to the progress achieved in terms of lower energy consumption, the use of enhanced materials for construction and improvement in the filter membranes' life [6]. Consequently, seawater desalination was quoted at around \$0.64-0.80/m³ in the mid-1990s, for a large size plant, operated based on a membrane process, compared to the quotation around \$.50/m³ for large-reverse osmosis plants.

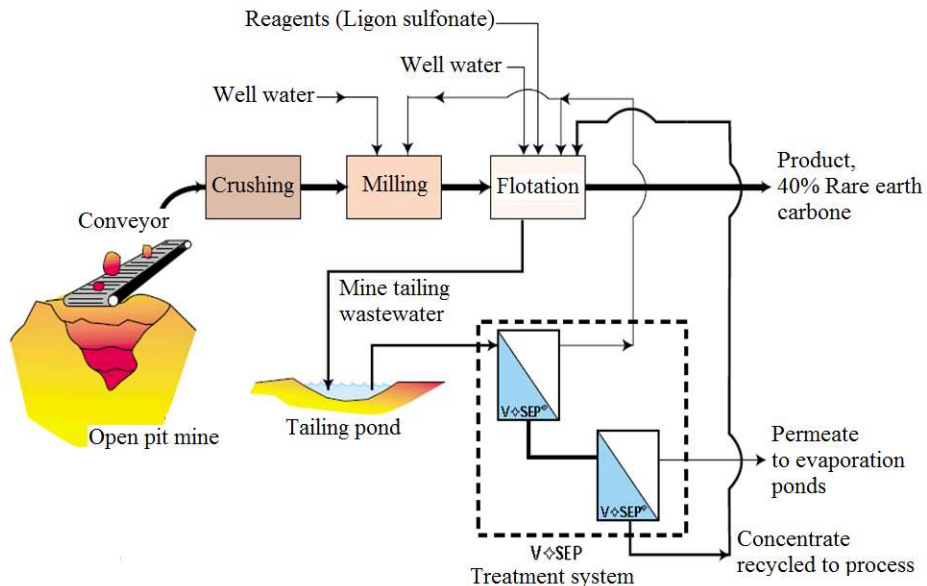


Figure 6: Water treated using membrane filtration and its recycling in ore flotation (after [37]).

Reverse osmosis utilizes nano-porous filter membranes (0.2-1.0 nm) and operates using dimensional exclusion or steric effect, electric charges, exclusion and solute-solvent-membrane physicochemical interactions to remove bacteria together with other pollutants present in the water as molecules and ions [6]. It is more effective than traditional techniques in removing bioorganic contaminants such as hormones and bioactive molecules of pharmaceuticals contained in wastewater emanating from households and biomedical centres [84].

Membrane technologies for wastewater reclamation have never been utilized in the Katanga province (DR Congo). They could play a leading role in the treatment of highly saline wastewater from the mining industry and thus help keep it under control and minimize adverse effects from the saltiness on aquatic wildlife of waterways utilized as spillways for industrial wastewater.

4.1.2. Pollutants Removal Through Flotation

Known as a mineral separation technique, flotation possesses interesting environmental applications as a method for treating industrial and household wastewater [24, 35, 87, 88]. In this area, it can be used in conventional mode as traditionally used in ores beneficiation [24] or as a dissolved air flotation (DAF) for treating wastewater (Table 9).

Table 9: Main differences conventional flotation and dissolved air flotation

Conventional Flotation	Dissolved Air Flotation
<ul style="list-style-type: none"> - Utilized in mineral ores beneficiation; - Solid-liquid separations (e.g., mineral and gangue); - Air dispersion in water (bulbs of 300-1500 μm); - Derive profit from the natural (talc, sulphides, coal, etc.) or induced wettability (use of surfactants and modifying agents, etc.) of minerals. 	<ul style="list-style-type: none"> - Utilized for wastewater treatment; - Solid-liquid or liquid-liquid separations (immiscible); - Micro bulbs formation (<100 μm); - Treatment of water contaminated by AMD and contaminants removal from water (oils, greases, colloids, hydrophobic ultrafine particles and dispersions, metals ion, etc.).

[89] have conducted at the laboratory scale conventional flotation in the removal of oils contained in domestic wastewater. A residual concentration slightly below the threshold value (50 mg/L) was achieved using aluminium sulphate as a coagulating agent at a pH of 5.5. In parallel, [90] showed that the combination of conventional flotation with reverse osmosis filtration allows the economical removal of copper contained in wastewater. Besides, high removal yields of copper were achieved using optimal doses of sodium dodecyl sulphate and polyacrylamide as a surfactant and a flocculent, respectively. DAF is currently utilized in the treatment of different types of industrial and domestic wastewater [24]. It enables the removal from the water of metal ions, oils as well as fine dispersions, micro particles, etc. The hourly treatment capacity of a flotation unit with dissolved air can reach 20,000 m³ of liquid effluents [91]. DAF is an emerging and expensive water treatment technique, which has been widely used in the treatment of wastewater from the oil industry, in the ores flotation (Figure 7) and in the coal mine in Brazil.

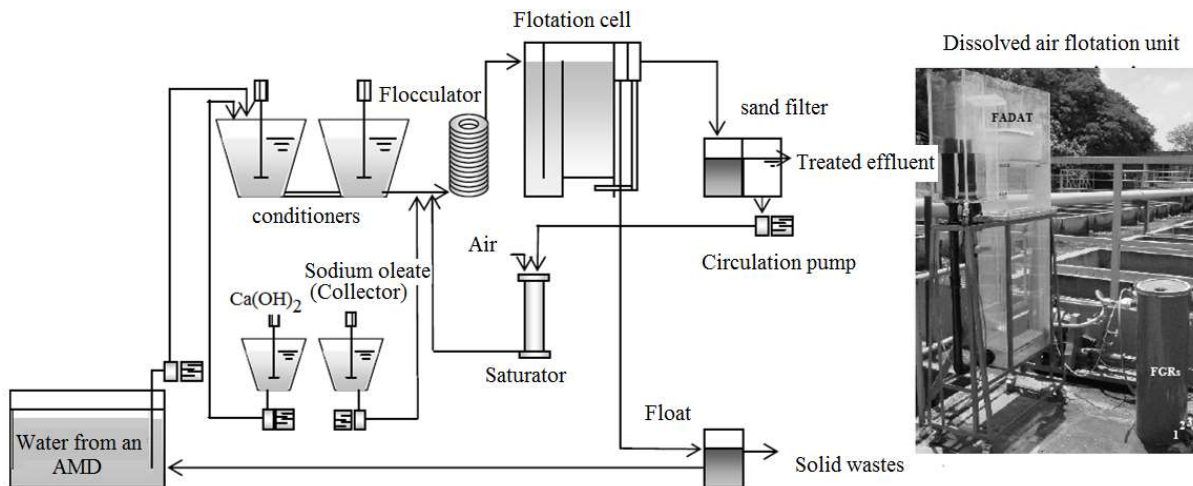


Figure 7: Decontamination of water polluted by an AMD using dissolved air flotation (After [38]).

In Katanga region (DR Congo), the Kakanda ores concentrator in view recycling wastewater from the flotation of copper-cobalt oxidized ores mined at the Mukondo deposit [92] had successfully utilized DAF in 2008. The widespread use of DAF in the treatment of wastewater from copper hydrometallurgical plants can help minimize the discharge of polluted waters into watercourses. It can also help produce both recyclable water intended for industrial processes and for less demanding applications from the quality standpoint and thus, promote the sustainable management of water resources in the Katanga region.

4.2. Physicochemical Treatment

Among conventional methods for treating inorganic waste waters such as discharges from the mining industry, adsorption [46, 93] has become one of the better treatment alternatives given that it can be implemented using various low-cost adsorbents and with larger metal-binding capacities [46, 94, 95]. Adsorption can be defined as a simple, useful and effective water treatment method [27, 96] that is based on a mass transfer process enabling to bind to a solid phase (an adsorbent) substances contained in water via physical or chemical interactions [93, 96]. According to [27], adsorption is acknowledged as one of the major treatment methods applicable to the decontamination of water considering that it is technically simpler and economically feasible [93, 96, 97]. Indeed, it is implemented using various adsorbents consisting of materials of mineral, organic and biological origin [27, 94]. The adsorbents can also consist of materials such as activated carbon, zeolites, as well as cheaper materials such as by-products from industrial processes, agricultural wastes and biosorbents consisting of stems, bark, leaves, including old newspaper, etc. [46, 93- 97]. [98] showed that adsorption is technically feasible to remove, with high efficiency (up to 92 % of chromium), heavy metal from industrial wastewater using composite biosorbent prepared by coating chitosan onto acid-treated oil palm shell charcoal, that is, a product derived from natural materials [46, 93]. However, according to [96], who were interested in zinc removal from wastewater, using adsorbents derived from waste material, adsorption has the disadvantage of failing to remove heavy metals when present to low concentrations [46, 93]. Moreover, its efficiency depends on the type of adsorbents used [96].

This wastewater treatment technique has never been used in the mining industry in Katanga region (DR Congo). However, tests carried out at the laboratory scale have proved the effectiveness of adsorption in the treatment of wastewater of industrial origin. For instance, liquid discharges (pH: 1.6; 86.99 mg/L Cu; 24.85 g/L Co and 16.47 mg/L Mn) released by the Shituru hydrometallurgical process have been treated through adsorption (20 minutes) of metal elements using the powder (1 g) of the plant species *Fiurena umbellatta* (particle size equal to 214 μm). Adsorption with this biosorbent type has enabled removing 71% of the copper, 63% of the cobalt and 27% of the manganese initially present in wastewater subjected to treatment. [99]. Another example to be mentioned is the treatment through adsorption using activated charcoal (with the particle size equal to 210-250 μm) of wastewater (pH: 8.34; 8.41 mg/L Cu; 7.35 mg/L. Co; 2.21mg/L Mn and 1.35 mg/L Fe) from the flotation of copper-cobalt ores. It enabled the removal of more than 99% of the majority of metal ions present in wastewater subjected to treatment [100]. Besides, the treatment through adsorption (10 minutes) of wastewater (pH: 1.65; 86.99 mg/L Cu; 24.85 mg/L Co and 22.57 mg/L Mn) from the Shituru Hydrometallurgical process, enabled removing 95% of the copper, 80% of the manganese and 70% of the cobalt. In this case, a biosorbent made of the biomass (0.5 g with the particle size of 202 μm) of the plant species *Juissaea abyssinica* was utilized [101].

4.3. Chemical Treatment

Chemical treatment of industrial liquid effluents in respect of environmental requirements is a common practice, worldwide [102]. Chemical treatment is an old method widely used for water polluted by an AMD [50, 103-106]. This water treatment method is based on the precipitation of dissolved contaminants by means of chemical reagents (CaO , Ca(OH)_2 , Na_2CO_3 , Mg(OH)_2 , NaHS , Na_2S , etc.), with the neutralization of its acidity [36]. Thus, chemical reagents destabilize dissolved or suspended contaminants in water by neutralizing their charges, forming bonds between particles to induce sedimentation [36, 103]. A chemical reagent can be used alone or mixed with a coagulant to increase treatment efficiency [103]. Some illustrations of the chemical treatment of wastewater are given below:

4.3.1. Neutralization with Lime

This technique is used in the treatment of wastewater from mining origin. It can both correct the pH of acidic waters and precipitate their pollutants (Table 10). According to [94], precipitation using lime is one of the most effective treatment means for inorganic liquid effluents of which metals concentration exceeds 1000 mg/L.

Lime is the most commonly used neutralizing agent (Table 10) in the treatment of polluted water. [107] have conducted at the laboratory scale the chemical treatment, using lime, of wastewater from the flotation of ores conducted at the Kipushi Concentrator (DR Congo). They have succeeded to rid wastewater of metal elements (Table 11).

Unlike other neutralizing reagents, lime is less expensive and provides heavy sludge with small volumes and without the possibility of remobilization of precipitated metal elements [103]. Its major drawback is the production of sludge with high pollution capacity. From the economic perspective and as with all precipitating agents, the reagents and sludge management's cost can be a major obstacle to the treatment of large volumes of wastewater using lime [50, 102]. That is the reason why the processes leading to high-density sludges are more preferred for rendering easy their management [44]. The treatment by lime of wastewater is a common practice at hydrometallurgical plants in Katanga region [57]. At the laboratory scale, there are numerous examples of the chemical treatment of industrial wastewater (pH: 7.82; 5.7 mg/L Co; 3.0 mg/L Mn; 118 mg/L Ca; 26 mg/L Mg; 0.2 mg/L Fe; 0.2 mg/L Ni; 1.8 mg/L Cu and 0.2 mg/L Pb) from the Katanga region (DR Congo) such as wastewater from the flotation of oxidized copper and cobalt ores. Among those examples, one can retain the treatment of wastewater from the ores flotation conducted at Kipushi through precipitation of metal elements using sodium carbonate (1.5 g/L) solution [108]. It had significantly enabled reducing the concentrations of cobalt (99.2%), manganese (92.3%), calcium (76.4%) and magnesium (88.2%). This same treatment had enabled reduction by half (50.0 %) the concentrations of iron and nickel with the elimination of more than half (57.1 %) of the copper and a quarter of the lead (25.2 %) present in the wastewater [108].

Table 10: Lime neutralization of mine water and copper flotation liquid effluents in the USA

Sample	Treatment pH	Total Metal Concentrations (mg/L)						
		Fe	Cu	Zn	Pb	Cd	As	Hg
Mine water & mill tailings*	6.5	0.35	1.09	22.8	0.01	0.32	0.006	0.0006
	7.0	0.06	0.33	5.4	<0.01	0.18	0.009	<0.0005
	8.0	0.05	0.06	0.29	<0.01	0.04	0.002	<0.0005
	9.0	0.05	0.04	0.09	0.01	0.02	0.004	<0.0005
	10.0	0.05	0.05	0.06	0.01	0.01	0.004	<0.0005
	11.0	0.10	0.04	0.04	<0.01	0.01	0.006	0.0008
Mill Tailings (Control)	-	0.06	0.09	0.06	0.01	0.01	0.003	<0.0005
		Dissolved Metal Concentrations (mg/L)						
Mine water & mill tailings*	6.5	0.06	1.00	22.1	<0.01	0.31	-	-
	7.0	0.03	0.26	5.4	<0.01	0.18	-	-
	8.0	0.02	0.06	0.29	<0.01	0.04	-	-
	9.0	0.02	0.04	0.09	<0.01	0.02	-	-
	10.0	0.02	0.05	0.06	<0.01	0.01	-	-
	11.0	0.02	0.03	0.04	<0.01	0.01	-	-
Mill tailings (control)	-	0.01	0.04	0.06	<0.01	<0.01	-	-

*Mine water: 5 parts; Mill tailings: 9 parts.

Table 11: Precipitation agents used in the treatment of polluted water

Product	Lime (Calcium Oxide or Hydroxide)	Soda Ash (Sodium Carbonate)	Caustic Soda (Sodium Hydroxide)	Magnesium Hydroxide
Form of material	Solid – CaO Powder – Ca(OH) ₂ Slurry 35% - Ca(OH) ₂	Power – Na ₂ CO ₃ Solution 15% - Na ₂ CO ₃	Solution 50% - NaOH	Slurry 58% - Mg(OH) ₂
Alkali requirement per ton H ₂ SO ₄ per ton HCl	As CaO: 1,240 lbs. 1,670 lbs.	2,190 lbs. 2,900 lbs.	1,630 lbs. 2,160 lbs.	1,190 lbs. 1,600 lbs.
Cost per ton of neutralizing agent (on a dry basis)	CaO – \$60 Ca(OH) ₂ – \$80 Slurry Ca(OH) ₂ - \$100 Cost stable	Na ₂ CO ₃ – \$80 Cost variable	NaOH – \$280 Cost highly variable	Mg(OH) ₂ – \$300 Cost increasing
Cost to neutralize 1 ton of H ₂ SO ₄	CaO – \$37 Ca(OH) ₂ - \$82	Na ₂ CO ₃ - \$86	NaOH - \$228	Mg(OH) ₂ - \$179
Maximum pH at 25°C	12.45	>11	14	10.6
Sludge profile	Heavy, low volume but easy handling even if heavy metals present	High volume, gel-like when heavy metals present	High volume, gel-like when heavy metals present	Heavy, low volume
Salts	Insoluble calcium metal Hydroxyl salts	Soluble sodium salts	Soluble sodium salts	Soluble magnesium Metal hydroxyl salts
TDS (total dissolved solids)	Low	High	High	High
Reaction time	Moderately fast-acting to complete neutralization	Moderately fast-acting to complete neutralization	Extremely fast-acting to complete neutralization	Fairly slow-acting to 95% complete neutralization

4.3.2. Neutralization and Purification Using a Reactive Drain

In this case, the chemical treatment is carried out via the passage of water inside a waterproofed underground trench constructed using a plastic film or a geomembrane. That underground trench was filled with grinded limestone, with the particle size comprised between 2-4 cm or 8-25 cm [44, 109]. It consists of a shallow (1-2 m) anoxic limestone-made drain (ALD) that is 0.6-1 m wide and a variable length of up to 20 m depending on the desired treatment degree and the flow rate of the water to be treated (Figure 8).

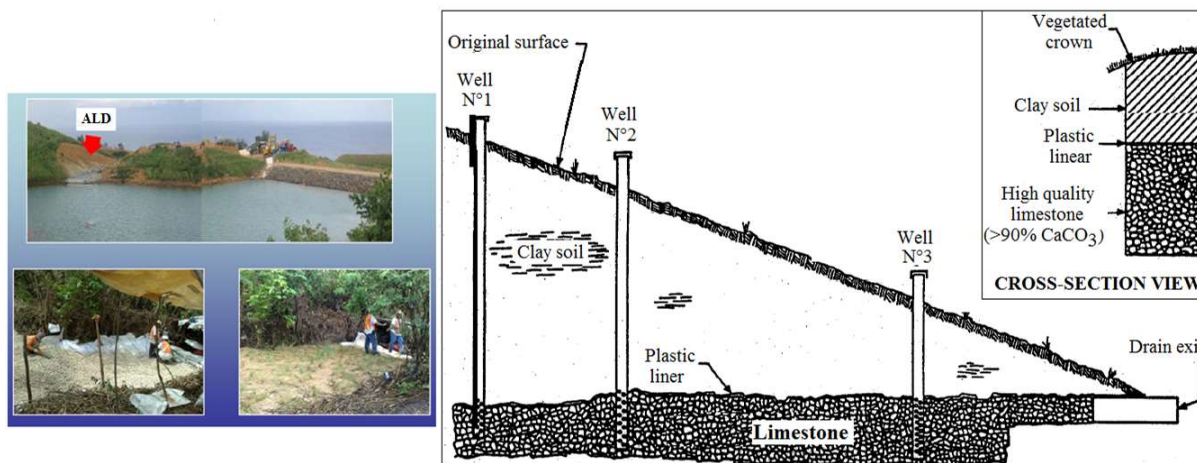


Figure 8: Treatment of AMD water using a limestone anoxic drain (ALD) (modified from [109]).

ALDs have water retention times ranging from 14 to 23 hours and are most commonly used for treating acidic wastewater containing metal elements and particularly those generated by AMDs [44, 109]. The treatment is based on the addition to water of calcium and bicarbonate ions through the dissolution of a high-grade limestone (82-99% CaCO_3 and 0.4-4.5% MgCO_3). This process brings about the increase in the pH of the water and the precipitation of metal elements. Around the neutrality ($\text{pH} = 7$), the removal of metal ions occurs through oxidation and hydrolysis, with precipitation happening downstream of the anoxic drain. That is why ALDs are always working in combination with wetlands in order to push further the treatment through oxidation and precipitation of iron including other water contaminant [44]. Anoxic environment that prevails inside the drain enables preventing limestone grains from undergoing passivation through the formation of a hydroxides layer. The same environment also prevents the water circulation channels from being obstructed [109]. The ALD performs passive, efficient and inexpensive treatment of wastewater [44] in spite of the fact that the process requires regular monitoring in view an efficient functioning.

For optimal water depollution using ALDs, the concentrations of dissolved oxygen, as well as those of iron and aluminium ions, must remain below 1 mg/L. As for the sulphate ion, its concentration must be less than 2000 mg/L. For low acidic or neutral polluted water [44], open limestone drains (OLD) are used to neutralize or buffer wastewaters before they pass through an aerated pond where the metal elements (Fe, Cu, Al, Zn, etc.) are removed through oxidation and precipitation. Despite the advantages that this treatment technique can offer in terms of the elimination of acidity and metal ions present in wastewater from mining sources, it has not yet been used in the Katanga region where cases of water and soil pollution by acidic muds have been recently experienced.

4.3.3. Precipitation of MTE Using Sulphide and Hydrogen Sulphide Ions

This wastewater treatment method utilizes the sulphides [Na_2S , $(\text{NH}_4)_2\text{S}$, NaHS , FeS , BaS or H_2S] as precipitating reagents of pollutants [44]. It has the advantage of producing sludge (precipitates) of base metal sulphides, which are less soluble than their hydroxides. Water pollution being usually more intense, the resulting sludge is chemically more stable or less reactive, due to a change in pH, when stored under anaerobic conditions [44]. Precipitation of metal elements using sodium sulphide is used in the treatment of wastewater from the flotation of Cu-Zn ores [104]. Given that the treatment is based on the neutralization of wastewater with lime, it also

produces highly polluting sludges. This treatment is used to clean up mining origin waters. It has been successfully implemented at the Raglan mine in Canada (Figure 9).

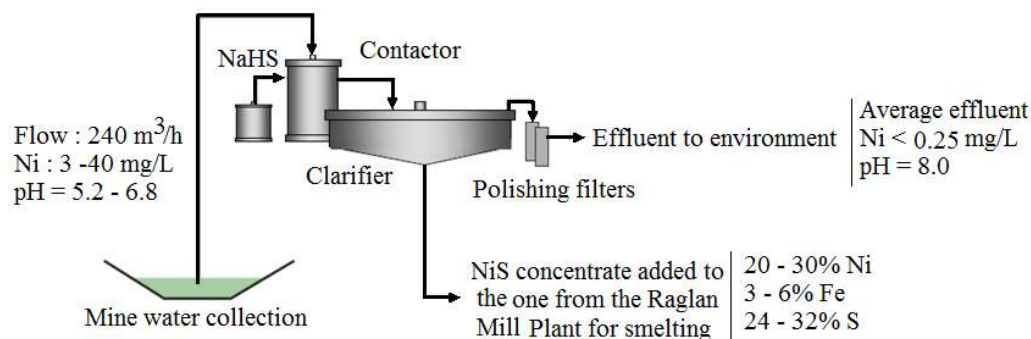


Figure 9: Precipitation of metals ions by means hydrogen sulphide ions (modified from [110]).

This treatment has the particularity of providing clean water of environmental quality as well as sludges that are rich in recoverable nickel sulphides [110]. However, the treatment units release harmful H₂S and produce colloidal precipitates difficult to separate from the treated water by means of ordinary filtration techniques. This renders very expensive the removal of pollutants from wastewater through precipitation of metal ions as sulphides comparatively to their removal as hydroxides in the presence of lime. Consequently, the treatment is not widely utilized in the clean-up of industrial wastewaters except in a few rare cases [44]. This type of water treatment has not yet been used at an industrial scale in the Katanga region (DR Congo). However, treatment tests conducted at the laboratory-scale on wastewater (pH: 8.04; 8.1 mg/L Mn; 0.2 mg/L Ni; 16 mg/L Mg; 1.4 mg/L Cu and 11.9 mg/L Co) from the flotation of copper and cobalt ores, using sodium sulphide as precipitating reagent, enabled successfully precipitating major metal elements [108]. Indeed, virtually all of the manganese (100,0 %), much of the nickel (95,1 %) and magnesium (98,2 %) as well as more than half of the copper (57.4 %) and cobalt (60.4%) initially present in wastewater were removed [108].

4.4. Biological Treatment

Biological treatment utilizes a passive process to clean up water [44, 102]. Although very demanding from the point of view of the implementation, it is more economical and environmentally friendly [26]. It does not generate contaminants and relies on the use of natural biological materials and processes. In the processing of mineral ores, biological treatment is used for ridding mining origin wastewater of pollutants for their safe discharge in watercourses or their recycling [22, 102]. The implementation of the biological treatment of wastewater is exemplified below.

4.4.1. Pollutants Digestion and Removal Using Biogenic Sulphides

The decontamination of polluted waters of mining origin is done using natural biodegradation or bacterial-stimulated processes [36]. A supply of nutrients or of sources of energy [102] such as lactates, pyruvates, formiates, malates, acetates and organic and inorganic pollutants themselves promotes the growth and breeding of bacteria involved in the pollutant removal processes. For illustrative purposes, *Bacillus polymyxa* was utilized in the biodegradation of surfactants [(CH₃)₂CHOCS₂Na, CH₃(CH₂)₁₁OSO₃NH₄, CH₃(CH₂)₇CH=CH(CH₂)₇COOH, etc.] used also as reagents in ores flotation [76, 110]. *Desulfovibrio* and *desulfotomaculum* have been used in the clean-up of polluted waters from AMDs through precipitation of metal elements using biogenic sulphides [102]. At Nickel Rim (Sudbury), the biological treatment was implemented using a Permeable Reactive Barrier (PRB) that enabled, removing iron (>30%) together with sulphate ions from water contaminated by an AMD [102]. A sulphate-reductive barrier (SRB) has been used by the Calliope Mine (Montana) in view of the removal of zinc (99%) from the water that has been in contact with acidogenic mining tailings [36, 102]. This passive method is presently considered the most promising alternative to traditional methods owing to its high treatment efficiency, low operating cost and strong reproducibility [68].

At Leviathan mine (California-USA), biological treatment of mining water (Table 12) has been conducted using a bioreactor inside which a bacterial reduction of sulphates has been implemented for metal elements precipitation as sulphides [112].

Table 12: Treatment of polluted water with sulphate reducing bacteria

Parameter	Treatment Duration			
pH	4.78	6.97	4.7	6.45
Temperature (°C)	9.9	15.6	8.4	13.1
Alkalinity (mg/L)	-	1,458	-	269
Al (mg/L)	41.0	0.02	48	0.24
As (mg/L)	0.41	0.023	0.28	0.015
Fe (mg/L)	310	2.8	380	260
Ni (mg/L)	1.8	0.01	2.1	0.01
Sulphate ion (mg/L)	1,690	1,190	2,070	1,910

4.4.2. Depollution of Water Using a Wetland

This passive technique for treating polluted water is based on biogeochemical reactions that occur in a natural aquatic environment [34, 44, 109, 113]. Wetlands are used in the treatment of industrial and domestic wastewater [114, 115]. They act as natural filtration media wherein take place biogeochemical processes [113-115] that are involved in the removal of pollutants (Figure 10).

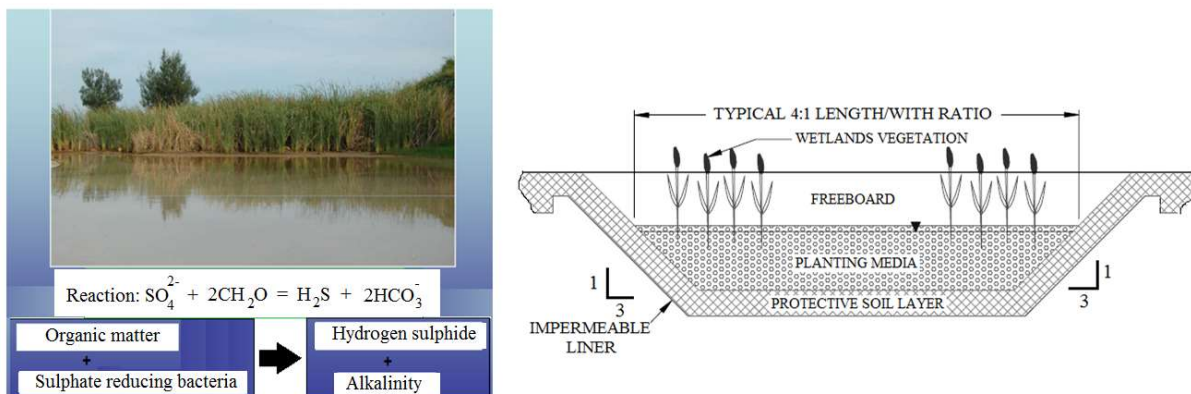


Figure 10: Depollution of water from an AMD using a wetland (modified from [34]).

Wetlands are usually swamps, shallow lakes, coastal areas, etc. Sometimes, they consist of areas flooded naturally or artificially by water with a certain frequency. Their emergent vegetation is most often made up of plant species such as *Typha latifolia*, *Phragmites australis*, *Eichhornia crassipes* and other macrophytes [26, 113, 115-120]. Wetlands constitute transitional ecological systems between land and the aquatic environment wherein the level of the water table is at or near the surface of the soil [121, 118, 119].

Wetlands operate on the principle of phytoremediation [26, 44, 122]. They consist of complex ecological systems where various chemical, physical and biochemical processes take place [114, 115] with a significant influence on water quality [109, 123, 124]. Wetlands are thus highly complex habitats where the interaction of biotic and abiotic factors results in a mechanical and biogeochemical filtration capable of removing pollutants from water [26, 115, 120]. The pollutant removal processes [26] and biogeochemical conditions that support their establishment in a wetland are described in Table 13 [34, 113].

Table 13: Biogeochemical processes for removing pollutants in a wetland

Treatment Processes	Biochemical Conditions in Constructed Wetlands	Constructed Wetlands Components		Examples of Constituents' Removal
		Hydro Soil	Plants	
Transfers				
Sorption	Availability and generation of surfaces	Low ratio of sand to clay; High organic matter content; porous	Large mass of roots and shoots	Hydrophobic chemicals (e.g., oil, grease, some pesticides, organometallics).
Volatilization	Presence of surface water and transpiration	Exposure to atmosphere (e.g., during drawdown)	Plants with transpiration rates	Chemical with high vapour pressure or low solubility; low-molecular-weight organics
Precipitation, settling and sedimentation	Flow rate less than approximately 10 cm/s conducive to settling (Stokes law)	Not applicable	Flow baffles to maintain flow rate and prevent short-circuiting	Solids and precipitates
Bioconcentration (plant uptake)	Prolific vegetation contact with the water	Favourable particle size and nutrients to support vegetative growth	Large mass in contact with water	Hydrophobic chemicals (e.g., oil, grease, some pesticides, organometallics)
Transformations				
Photolysis	Sunlight intensity and light absorption	Not applicable	Minimize shading	low solubility; low-molecular-weight organics
Hydrolysis	Acid, basic or neutral environment depending on targeted constituents	Not applicable	Not applicable	Pesticides
Speciation and ionization	Presence of reactive ions or electrons (e.g., oxidation, reduction)	Refer to oxidation and reduction (below)	Refer to oxidation and reduction (below)	Metals and organics
Oxidation	Redox (Eh) > -50 (approximately); pH slightly acidic to near neutral	High rate of sand to clay; Low organic matter content	Rhizosphere aeration; large radial oxygen loss	Organics (e.g., oil and grease); some metals (e.g., Fe)
Reduction	Redox (Eh) < -150 (approximately) pH near neutral to slightly basic	Low rate of sand to clay; High organic matter content	small radial oxygen loss; root metabolism in anaerobic environment	Metals (e.g., Hg, Cu, Pb, Zn); organochloride chemicals subjected to dehalogenation
Biotransformation and biodegradation	Presence of organisms and enzymes capable of transforming targeted constituents	Favourable particle size and nutrients to support microbial growth	Plant to support periphytic and Rhizosphere microbial growth	Biodegradable organics

The environment that prevails inside a wetland [34, 117] dictates the establishment and the extent of biogeochemical processes implicated in the depollution of water, including oxidation, reduction, chelation, adsorption, complexation, sedimentation, filtration, biosorption of pollutants by plants and their microbial digestion. The hydrological regime, geology and sediment chemistry, the composition of the microbial community, fauna and flora, and the geomorphology and position of the wetland in a given landscape [121] influence this environment. Thus, in a wetland, oxidizing conditions prevail on the water surface up to the water-sediment interface [115] and this is not the case in the depth of the sediments where reducing (anaerobic) conditions predominate [113,115]. The bacteria (sulphate-reducing species, denitrifying species, etc.) that live in wetlands catalyse the retention reactions of metal elements [44, 119]. As for metal elements present in solution found in the oxidation zone, they are eliminated through biosorption by macrophytes, which constitute a natural filter

wherein oxidation processes take place [44, 117]. Metal elements are also removed through chelation or via the organic complexation reactions [123].

In the anaerobic zone of a wetland, bacteria reduce sulphates to hydrogen sulphide through the oxidation of organic matter [44]. Metal ions (Zn^{2+} , Pb^{2+} , Cu^{2+} , and other metal ions) are retained by precipitation (insoluble compounds) as the outcome of their reaction with hydrogen sulphide (Figure 10). For a system that has already reached its sorption limit, the most dominant metals removal mechanism becomes sulphide formation [123].

In a wetland, an anoxic or reducing environment results from microbial or animal respiration together with the slowed diffusion of oxygen into water and deep sediments [119]. Depending on the nature of the pollutants targeted, it is, therefore, possible to promote the establishment of an oxidative or reducing process, that is to say, the maintenance in a wetland of aerobic or anaerobic conditions that favour a given reactional process [109].

Since 1950, wetlands have been utilized successfully in the treatment of nutrient-rich runoff intended for agricultural use [113, 115]. They are presently utilized in the treatment of contaminated waters from the industrial origin, polluted and acidic water generated by AMDs, municipal wastewater and petroleum refining wastewater, including organic and inorganic wastewater from landfills or those generated due to leaching of solid tailings of the mining and metallurgical processing of ores [34, 44, 109, 113-115, 120,123,124]. Wetlands are also utilized to control the water runoff pollution in urban areas, including the pollution of storm-generated runoff [113, 117, 121, 122].

In Canada, the decontamination of AMDs water had been conducted using passive treatment techniques such as ALDs, biosorption-based systems and wetlands [109, 114]. Indeed, in a wetland, the contaminated water from AMDs can flow by gravity and undergo both depollution and neutralization, respectively, via the reactions of precipitation, chelation and ionic exchange as well as via the action of the sulphate-reducing bacteria [114].

Currently, wetland performance (Table 14) makes them more attractive in terms of efficiency and cost compared to traditional wastewater treatment techniques such as lime neutralization, the process based on activated sludge or percolation filtration [115, 117-120, 123].

Table 14: Treatment of AMDs from coal mines using wetlands in England

Wetland	Mine Water Total Iron Loading (kg/day)	Outlet Total Iron Loading (kg/day)	Removal Rate (%)	Removal Rate (g/day/m ²)	Acidity (As Free CO ₂) Removal Rate (g/day/m ²)
A	9.3	1.5	83.2	8.1	9.5
B	52.8	2.6	95.3	7.9	15.4
C	38.8	1.1	96.0	5.8	9.3

In addition, a wetland is more suitable for the treatment of large volumes of polluted water [34, 118]. A relevant illustration of this passive technique is given by the treatment of mining waters and wastewater generated by the flotation of Pb-Zn ores in the Philippines. Thus, there are several benefits to derive from the use of wetlands in the treatment of polluted water [117, 123, 124]. These include, but are not limited to the following:

- simpler construction with local materials and at low cost;
- adaptability to the quantity rather than to the quality of the water to be treated;
- stable operation based on natural processes for pollutant removal;
- easy implementation and maintenance;
- great treatment efficiency and affordability.

Concerning mining wastewaters, the combination of other treatment techniques such as adsorption or a system using a limestone trench in combination with constructed wetlands [44] has proven successful in the

removal of heavy metals. Indeed, [125] have succeeded in achieving the removal of heavy metals (As-80.3%, Mn-96.9%, Cd-79.6%, Zn-52.9% and Pb-38.7%) contained in mining wastewater after four months of treatment. The wastewaters were emanating from a Pb-Zn mine located in northern Vietnam, with a constant flow rate of 5m³/day, and have been treated by adsorption. Modified iron-ore drainage sludge, in combination with a horizontal-subsurface-flow constructed wetland made with common reed (*Phragmites australis*) was utilized for wastewater treatment. Interested in improving the management of liquid effluents from the copper ore flotation in the Katanga region (DR Congo), [36] have considered the possibility of treatment of wastewater by biological techniques. They have suggested utilizing a wetland as a polishing stage after ridding wastewater of pollutants through treatment inside a tailings pond converted into a bioreactor.

5. Conclusion

Water plays a very essential role in human life and in the prosperity of all industries. However, water is an unevenly distributed resource on planet Earth. The water crisis includes the increasing deterioration in its quality and its ever-growing demand in the face of limited access to the resources available in the world. Accordingly, the development of water reuse, wastewater treatment processes and remediation systems has gained importance in view of increasing access to water. Different technologies, based on the utilization of physical, chemical, physicochemical and biological processes either alone or in combination with others in view to rid the wastewaters of pollutants, are utilised in the treatment of raw water and wastewaters of different origins, including those of the mining industry. They enable recovering water that can be reused either in other sectors or safely released to watercourses. It is important to keep in mind that each treatment technology has both advantages and disadvantages on efficiency, the implementation and maintenance cost, the operational time and technique as well as space requirements. These different criteria, together with others not addressed in the present work, are used to determine which treatment technique can be selected for the treatment of wastewater of mining origin.

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